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# Comparison of Pt-BaO/Al<sub>2</sub>O<sub>3</sub> and Pt-CeO<sub>2</sub>/Al<sub>2</sub>O<sub>3</sub> for NOx storage and reduction: Impact of cycling frequency



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#### ABSTRACT

NOx reduction under net lean and near-stoichiometric conditions was carried out on  $Pt/Al_2O_3$ ,  $Pt/CeO_2/Al_2O_3$  and  $Pt/BaO/Al_2O_3$  washcoated monoliths to compare performance features and identify reaction pathways. The impact of the storage components (BaO, CeO<sub>2</sub>) on the NOx conversion and byproduct (NH<sub>3</sub> and N<sub>2</sub>O) yields was quantified for a range of feed temperatures, reductant types (H<sub>2</sub>, and  $C_3H_6$ ), O<sub>2</sub> feed concentrations, and cycle times. The NOx storage functionality is essential for NOx reduction under net lean conditions while the oxygen storage functionality promotes NOx reduction for near-stoichiometric conditions. NOx conversion by H<sub>2</sub> under lean conditions is dependent on the NOx storage capacity of the catalyst, with  $Pt/CeO_2/Al_2O_3$  and  $Pt/BaO/Al_2O_3$  exhibiting the highest NOx conversion below and above 300 °C, respectively. High NOx conversion is achieved over  $Pt/CeO_2/Al_2O_3$  for anaerobic rich feeds at temperatures above 400 °C. Increasing the O<sub>2</sub> feed concentration enhances NOx conversion over  $Pt/CeO_2/Al_2O_3$  below 400 °C but inhibits NOx conversion above 400 °C. The former is attributed to promotion of NO oxidation leading to NOx storage while the latter is attributed to O<sub>2</sub> inhibition of NO decomposition/reduction. Shorter cycle times increase the NOx conversion with  $C_3H_6$  as reductant over  $Pt/BaO/Al_2O_3$  under lean conditions and over PCA for the near-stoichiometric feed. The findings confirm that improved NOx storage utilization is mainly responsible for NOx conversion enhancement. A ceria redox pathway has only a secondary effect on NOx conversion under excess O<sub>2</sub>.

# 1. Introduction

Several technologies developed for NOx abatement are under increasingly stringent regulations [1]. NOx storage and reduction (NSR) [2] and selective catalytic reduction (SCR) [3] are the most effective but have limitations. Notably, achieving adequate NOx conversion either at low or high temperatures (< 200 °C or > 450 °C) remains elusive as does the generation and/or release of undesired species such as N2O and NH3. NSR [4] involves cycling of lean and rich feeds to the lean NOx trap (LNT). The rich feed is obtained through modulation of the air-fuel ratio or by fuel injection into the exhaust. The rather narrow temperature window giving high NOx conversion is ascribed to kinetic limitations at low feed temperature and to NOx storage capacity, masstransport limitations and exothermic heat effects at high feed temperatures [4,5]. In contrast, SCR utilizes NH3 obtained by decomposition of injected aqueous urea for selective reduction of NOx under net lean conditions. The inadequate NOx conversion at low (< 200 °C) and high temperatures (> 450 °C) [2,6] is due to constraints of the urea feed system and to the undesired oxidation of ammonia, respectively.

The limitations of NSR and SCR define potential approaches for

improved performance. Several advanced deNOx techniques or designs have been proposed, such as combination of NSR and SCR [7–9], novel cold start catalysts [10,11], SCR catalyst coated on soot filters, high porous flow substrates [12,13], among others. About 10 years ago Toyota researchers [14] introduced the Di-Air (diesel NOx aftertreatment by adsorbed intermediate reductants) technology, a variant on NSR that involves fast injection of fuel (reductant) into the exhaust upstream of a NSR converter. The Di-Air system is superior to conventional NSR system in many aspects, particularly at high temperatures and space velocities, typically encountered during high load vehicle operation [15].

The enhanced NOx conversion achieved in the Di-Air system has led to follow-up studies focused on mechanistic investigations and system design and optimization. Zheng et al. [40,41] investigated the enhanced deNOx performance over a dual-layer LNT + SCR catalyst. Uenishi et al. [42] and Reihani et al. [43] proposed improved designs to minimize axial dispersion upstream of the NSR converter which enabled more intense hydrocarbon pulses. Faster cycling generally has led to higher NOx conversion with hydrocarbons typically outperforming  $\rm H_2$  or CO at high temperature (e.g.  $> 350\,^{\circ}\rm C$ ). Both Reihani et al. [18]

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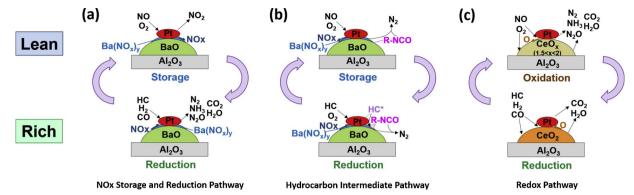


Fig. 1. Three working mechanism possibly involved in the Di-Air system: (a) NOx storage and reduction mechanism; (b) hydrocarbon intermediate mechanism; (c) redox pathway. Here BaO and CeO<sub>2</sub> are chosen to represent the NOx and oxygen storage material respectively.

and Li et al. [19] pointed out that olefins (e.g.  $C_3H_6$  or  $C_2H_4$ ) are better reductants than the corresponding alkanes. Three mechanisms depicted in Fig. 1 have been proposed to explain the enhanced conversion with fast cycling NSR. These include (a) hydrocarbon intermediates pathway, (b) enhanced NOx storage utilization, and (c) and ceria redox.

The hydrocarbon intermediate pathway (Fig. 1b) proposed [21] by Toyota researchers was based on a combination of reactivity and surface spectroscopy data [14–16]. Isotopic labelling experiments revealed the existence and consumption of thermally stable hydrocarbon intermediates such as nitrous acid esters and amides on the catalyst surface during fast cycling [15]. They concluded that such intermediates, which survive on the surface during fast cycling at elevated temperatures, are subsequently oxidized selectively to  $N_2$ . This mechanism was claimed to be mainly responsible for the enhanced NOx abatement performance in the Di-Air system [15].

The conventional NSR pathway (Fig. 1a) with enhanced NOx storage utilization has been proposed and examined in several recent studies [20,22,23]. Reihani et al. [22] compared the dynamic NOx storage capacity at low and fast cycling frequencies and concluded that fast cycling results in more efficient NOx storage and reduction and therefore higher NOx conversion. Ting et al. [20,24] confirmed the presence of hydrocarbon intermediate pathway but ascribed the majority of the enhancement by fast cycling to the conventional NSR mechanism through comparison of the overall deNOx performance with  $\rm H_2$  or  $\rm C_3H_6$  as the reductant. Ting et al. [23] used a non-isothermal 1-D monolith reactor model containing a global kinetic scheme to capture the experimental trends in  $\rm H_2$  case based on the conventional NSR pathway.

The cyclic redox of ceria has been advanced by Makkee and coworkers [25-28,38,74,75]. They used TAP (Temporal Analysis of Products) to show that oxygen vacancies are generated at high temperatures on reduced ceria which serve as sites for NO decomposition/ reduction. Ceria has been widely studied since its first application by Ford [29] in 1970s. Due to its unique capability of oxygen incorporation into its crystal lattice [29,30], ceria has been added to supported metal and metal oxide catalysts for enhanced oxidation of emitted reductants, such as the three-way catalyst (TWC) [29], diesel oxidation catalyst (DOC) [31] and diesel particle filters (DPF) [32]. In addition, ceria adsorbs NOx at low temperature, motivating its use in NSR [33-35] and passive NOx adsorption (PNA) [36,37]. Wang et al. [25–28] further showed that carbonaceous species deposited during the pulsing of the reductant (H2, CO, HC) were converted under the subsequent oxidizing (NO, O2, CO2) conditions. While a supported ceria catalyst was effective at high temperatures, the addition of precious metal lowered the temperature needed for the reaction (Pt/CZO and Rh/CZO). Our recent study [39] showed that while NO is effectively converted on reduced ceria at high temperature (> 450 °C), the reduction effectiveness is largely limited due to the competing oxidation by excess  $O_2$  in the feed.

Among the three mechanisms, the conventional NSR pathway and the hydrocarbon intermediates pathway require the participation of adsorbed NOx, accomplished by NOx storage material (e.g. BaO, CeO<sub>2</sub> at low temperature). In contrast, the redox pathway is realized by incorporation of oxygen storage material (e.g. CeO<sub>2</sub>).

This study investigates the impact of oxygen and NOx storage materials (CeO $_2$ , BaO) on NOx conversion during fast lean-rich switching. The deNOx performance of three catalysts, Pt/Al $_2$ O $_3$ , Pt/CeO $_2$ /Al $_2$ O $_3$ , Pt/BaO/Al $_2$ O $_3$ , are systematically compared over a range of operating conditions, including cycling frequency, feed temperature, reductant type (H $_2$ , C $_3$ H $_6$ ), and oxidant (O $_2$ , CO $_2$ , H $_2$ O) concentrations. The findings are interpreted to advance the understanding and application of fast cycling NSR.

# 2. Experimental

## 2.1. Catalyst preparation

Three washcoated monolith samples, Pt/Al<sub>2</sub>O<sub>3</sub> (hereafter denoted as "PA"), Pt/CeO<sub>2</sub>/Al<sub>2</sub>O<sub>3</sub> ("PCA") and Pt/BaO/Al<sub>2</sub>O<sub>3</sub> ("PBA") were prepared first by synthesizing the powders and then coating small monoliths by slurries containing the active materials. The total washcoat loading was 4.6 g/in<sup>3</sup> and the Pt loading was 90 g/ft<sup>3</sup>. All three catalyst contained ~1 wt. % Pt (washcoat) while the PCA and PBA contained 11 wt. % CeO2 and BaO, respectively. The powder catalysts were prepared by the incipient wetness impregnation method. The  $\gamma$ -Al<sub>2</sub>O<sub>3</sub> (Sasol Inc.) support was calcined in static air at 550 °C for 2 h before impregnation. For PCA or PBA, the γ-Al<sub>2</sub>O<sub>3</sub> powder was impregnated with aqueous solution of cerium (III) nitrate hexahydrate (Sigma-Aldrich) or barium nitrate (Sigma-Aldrich), respectively. The powders impregnated with Ce or Ba precursors were dried overnight at 120 °C and calcined in static air at 550 °C for 2 h. The Pt impregnation was achieved by impregnating CeO<sub>2</sub>/Al<sub>2</sub>O<sub>3</sub>, BaO/Al<sub>2</sub>O<sub>3</sub> and γ-Al<sub>2</sub>O<sub>3</sub> powders with aqueous solution of tetraammineplatinum (III) nitrate hydrate (Sigma-Aldrich). The powders impregnated with Pt precursor were dried overnight at 120 °C and calcined in static air at 550 °C for

The Pt/CeO $_2$ /Al $_2$ O $_3$ , Pt/BaO/Al $_2$ O $_3$  and Pt/Al $_2$ O $_3$  particles were deposited onto cordierite monolith pieces by a dip-coating method. Blank cordierite monoliths (Mitsubishi, 400 channels per square inch, 1 inch diameter, 2.5 inch in length) were cut into small pieces of (0.42 inch) diameter and (1 inch) length, having ~56 channels. To prepare the PA monolith sample, the PA powder was first ball-milled along with boehmite (AlOOH; Nyacol Nano Technology Inc.) contained in deionized water for 20 h. The monolith piece was immersed from both ends into the Pt/Al $_2$ O $_3$  slurry for 30 s. Air was then blown for 15 s from both ends of the monolith to remove excess slurry and get a uniform coating. After each round of dip-coating, the monolith was dried at 120 °C for 2 h. Successive dip-coating was applied until the desired

**Table 1**Feed gas compositions for cyclic experiments.

Catalyst	BET Surface Area (m <sup>2</sup> /g)	Pt Dispersion (%)	
PA	119.9	5.3	
PBA	99.7	3.4	
PCA	112.1	12.4	

washcoat loading was achieved (4.6 g/in³). The dip-coated monolith pieces were calcined at 550 °C for 2 h. Monolithic PBA and PCA samples were prepared by a similar dip-coating method, the three monolithic catalysts were aged before reaction evaluation in static air at 700 °C for 33 h, which is equivalent to  $160,000\,\mathrm{km}$  aging [14].

 $\rm N_2$  physisorption and CO chemisorption tests were conducted to determine the surface area and Pt dispersion for the three samples using a Micromeritics 3 Flex instrument. Catalyst washcoat was scraped from monolith samples for characterization. The results are summarized in Table 1. Physisorption for surface area was carried out using  $\rm N_2$  at 77 K. Chemisorption for Pt dispersion was determined using CO at 35 °C. Physisorption tests show that the three samples (PA, PBA and PCA) had similar. That each of the dispersions is less than 15% is likely a result of the accelerated aging protocol employed. The Pt dispersion for PCA is higher than the values for PA and PBA. This may be due in part to additional CO consumption by reduction of ceria and adsorption on ceria.

SEM was used to determine the thickness of washcoat layer. The thickness of washcoat varied peripherally with a thicker thickness in the corner and a thinner thickness on the wall. The peripherally-averaged thickness for PA, PBA and PCA are  $39\,\mu m,\,43\,\mu m,\,41\,\mu m$  respectively. Weisz-Prater modulus was calculated to assess the impact of internal diffusion. Detailed discussion was included in Supplementary material (section 7). The washcoat diffusion cannot be isolated in this study but the diffusion impacts in the three samples would be similar, as the three samples has similar washcoat thickness and porosity.

# 2.2. Bench-scale reactor set-up

A detailed description of the bench-scale flow reactor setup is described in previous studies [40,41]. The feed gas mixture was supplied by a bank of gas cylinders (Matheson Tri-Gas; Praxair, Inc.) and metered by mass flow controllers (MKS Inc.). The gases were fed to the reactor via three different lines: lean line (NO, O2, Ar), rich line (NO, H<sub>2</sub>, C<sub>3</sub>H<sub>6</sub>, Ar), and main line (CO<sub>2</sub>, H<sub>2</sub>O, Ar). The main line feed was connected to enable mixing with either the lean or rich line feed upstream of the catalyst. A solenoid-actuated four-way valve (Valco Inc., Micro-electric two position valve) enabled shifting between lean and rich feed at a prescribed switching frequency. A syringe pump (Teledyne Isco model 100DX) and vaporization system were utilized to inject vaporized H2O into the feed gas. The distance between the injection point and the monolith was ~2 ft in order to reduce the undesired axial mixing (dispersion) of the lean and rich streams, to enable a better comparison to data from previous studies [19,39,40]. The monolithic sample, wrapped by a Fiberfrax® ceramic paper, was placed in a quartz tube (OD = 0.75 in.), contained in a furnace (Thermocraft). One Ktype stainless steel sheathed thermocouple (Omega Engineering Inc.) was placed at the geometric midpoint of the monolithic sample. A FTIR (Thermo Scientific, Nicolet 6700) monitored the concentration of NO, NO2, N2O, NH3, CO, CO2, C3H6 and H2O. A mass spectrometer (Hiden Analytical, HPR20) monitored the  $N_2$  concentration (m/e = 28). The mass spectrometer was calibrated daily before the start of experiments. When H<sub>2</sub> was used as the reductant and CO<sub>2</sub> was absent in the feed, the N balance was closed within 5%. Otherwise, the amount of N2 generated was calculated from the overall nitrogen balance because of the overlap of m/e = 28 signal from CO and  $CO_2$  in the mass spectrometer. When C<sub>3</sub>H<sub>6</sub> was used as the reductant, the C balance is closed within

 Table 2

 Feed gas compositions for cyclic experiments.

Run	Rich		Lean	Lean	
	Reductant	NO/ ppm	O <sub>2</sub> /%	NO/ppm	
1	6.21% H <sub>2</sub>	300	5	300	9.7
2	6.21% H <sub>2</sub>	300	0.5	300	1.0
3	6900ppm C <sub>3</sub> H <sub>6</sub>	300	5	300	9.7
4	$6900$ ppm $C_3H_6$	300	0.5	300	1.0

5%. When  $C_3H_6$  was the reductant, the N balance cannot be assessed because of the aforementioned m/e=28 overlap. For both the  $H_2$  and  $C_3H_6$  reductant experiments, H balance and O balance cannot be calculated, as  $H_2O$  signal monitored by FTIR is not accurate enough for quantitative evaluation.

#### 2.3. Lean-rich cyclic experiments

Lean-rich cycling experiments were conducted over the three monolithic catalysts (PA, PCA and PBA). The data for the PBA and PCA samples are compared to PA to assess the contribution of the BaO and CeO<sub>2</sub>, respectively. Two reductants (H<sub>2</sub> and C<sub>3</sub>H<sub>6</sub>) tested the overall NOx abatement performance obtained by the three catalysts. The total system flowrate was 3000 sccm, corresponding to a gas hourly space velocity (GHSV) of  $76,000\,h^{-1}$  (at STP). Table 2 summarizes the feed conditions used in the cyclic experiments. The lean to rich time ratio was 6 to 1 (14.3% rich duty cycle). The fixed duty cycle set the fuel to oxidant ratio. The lean/rich frequencies used in the study spanned 90 s/ 15 s,  $60\,s/10\,s$ ,  $30\,s/5\,s$  and  $6\,s/1\,s$ . The cycle-averaged stoichiometric number was used to characterize feed stoichiometry by

$$S_N = \frac{[NO] + 2[O_2]}{[H_2] + 9[C_3H_6]} \tag{1}$$

The  $\rm O_2$  concentration was varied to provide a cycle-averaged stoichiometric feed, either ( $\rm S_N=1.0$ ) or lean feed (i.e.  $\rm S_N=9.7$ ). In some experiments the  $\rm O_2$  concentration was varied to study its impact on the deNOx performance (i.e.  $\rm S_N$  varying from 0.03 to 9.7). In most experiments,  $\rm CO_2$  and  $\rm H_2O$  were excluded to minimize the impact from side reactions (i.e. water gas shift reaction and competitive adsorption on BaO or  $\rm CeO_2$  sites).  $\rm CO_2$  and  $\rm H_2O$  were added in several experiments to study the impact of  $\rm CO_2$  and  $\rm H_2O$ .

The cyclic experiments were carried out at feed temperatures ranging from 150 °C to 600 °C with 50 °C temperature increments. Before each set of experiments, the catalyst was pretreated in 2%  $\rm H_2$  until NH $_3$  was not detected in the effluent, then by 2%  $\rm O_2$  for 10 min, and finally by a 10 min Ar purge. This procedure ensured that no stored NOx accumulated and that the catalysts were fully oxidized. Once a cyclic stationary-state was reached, results from up to 10 cycles were averaged to determine the cycle-averaged reactant (NOx and  $\rm C_3H_6$ ) conversion and product (N $_2\rm O$  and NH $_3$ ) yield. The cycle-averaged reactant (NOx and  $\rm C_3H_6$ ) conversions ( $\rm X_{NO_X}$ ,  $\rm X_{C_3H_6}$ ) and NH $_3$  and N $_2\rm O$  yields ( $\rm Y_{NH_3}$ ,  $\rm Y_{N_2O}$ ) were calculated by standard expressions reported in the Supplementary Material.

In the current study our experimental approach was to turn off the  $O_2$  during the rich injection in order to avoid the steep exotherm [17,20]. For example, Ting et al. [20] reported temperature rise of up to  $300\,^\circ\text{C}$  during propylene injection into an  $O_2$ -containing feed, under net rich conditions. This approach enables a clearer performance comparison of the Pt/BaO/Al $_2O_3$  and Pt/CeO $_2$ /Al $_2O_3$  catalysts, which is the main objective of this study. In one of the first Di-Air reports by Toyota researchers, Inoue et al. [14] showed that high NOx conversion may be achieved through rapid reductant injection into a simulated lean exhaust stream while maintaining an overall lean condition throughout the cycle. However, other researchers reported that high NOx conversion could only be achieved with a net rich rapid pulse [17,19,20,43].

In our studies, enhanced NOx conversion could be achieved with a net rich pulse.

# 2.4. NOx storage experiments

The NOx storage capacities of each of the three samples (PA, PCA and PBA) were measured under a continuous feed of 300 ppm NO and 5%  $\rm O_2$ . Prior to each NOx storage experiment, the catalyst was fully regenerated by applying the pretreatment procedure described above. The NOx storage capacity,  $C_{NO_x}$ , is calculated from

$$C_{NO_X} = \frac{F_{NO_X} t^* - \int_0^{t^*} F_{NO_X}(t) dt}{m_{wc}}$$
 (1)

where  $t^*$  is the duration and  $m_{wc}$  the washcoat mass.

#### 2.5. Steady-state experiments

Selected steady-state experiments were carried out to determine the impact of mixing over the PCA sample using feeds with various  $S_{\rm N}$  values. In all steady-state experiments, the feed temperature was 550 °C and contained 300 ppm NO, 8891 ppm  $\rm H_2$ , varied concentration of  $\rm O_2$  and balance Ar. The  $S_{\rm N}$  was varied between 0.5 and 1.5 through variation of the  $\rm O_2$  concentration. The PCA sample was pretreated following the procedure described in 2.3.

#### 3. Results

# 3.1. Impact of catalyst formulation

In this section, we compare the overall NOx abatement performance over each of the catalysts for a cycle-averaged lean feed ( $S_N = 9.7$ ) conducted at the fastest cycling frequency (6/1 s) and using with  $H_2$  and  $C_3H_6$  as reductants.  $H_2O$  or  $CO_2$  were excluded in these experiments to avoid interference from side reactions. The impacts of  $H_2O$  and  $CO_2$  are primarily discussed in the Supplementary Material (section 6) [20].

Fig. 2 compares the cycle-averaged NOx conversion for the three catalysts as a function of feed temperature with either H<sub>2</sub> (2.a) or C<sub>3</sub>H<sub>6</sub> (2.b) reductant. For both H2 and C3H6, PA is inferior to PBA and PCA over the entire temperature range. This confirms the promotional impact of BaO in storing NOx and of CeO<sub>2</sub> in storing O<sub>2</sub> and NOx. For H<sub>2</sub>, the NOx conversion using PA or PCA decreases monotonically with feed temperature while for PBA the conversion exhibits a local maximum at ~300 °C. PCA outperforms PBA in the lower temperature range (T < 300 °C) while PBA outperforms PCA for intermediate and high temperatures (T > 300 °C). For C<sub>3</sub>H<sub>6</sub>, the NOx conversion obtained with both PBA and PCA exhibits a local maximum at an intermediate temperature; specifically ~50% at ~400 °C for PCA, and ~65% at ~450 °C for PBA. In contrast, the NOx conversion obtained with PA is below 10% over the entire feed temperature range. The NOx conversion for PCA slightly exceeds that for PBA for feed temperatures below 400 °C while PBA significantly outperforms PCA at higher temperatures. These trends are consistent with those reported by Yang et al. [41] who showed that the addition of ceria to Pt/BaO/Al2O3 promotes cycleaveraged NOx conversion over a wide intermediate temperature range (175 °C-400 °C). The promotional impact diminishes at high temperatures (~4% at 400 °C). The specific differences in the NOx conversions for the three catalysts can be largely explained by their discrepant NOx storage capacities. We return to this point below.

Fig. 2.c shows the cycle-averaged  $C_3H_6$  conversion over each of the three catalysts as a function of feed temperature. (Note: The  $H_2$  conversion in the  $H_2$  experiments is not reported as the  $H_2$  was not measured.) For each catalyst the  $C_3H_6$  conversion remains low (< 15%) for temperatures below 300 °C. The merging of the conversions in this range is reflective of the equal Pt loading (1 wt.%). Upon light-off, the  $C_3H_6$  conversion rank is PCA > PBA > PA. Propylene consumption

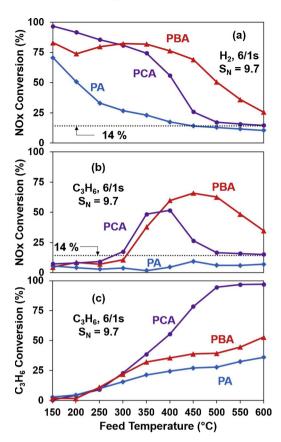


Fig. 2. Cycle-averaged reactant (NOx and  $C_3H_6$ ) conversions as a function of feed temperature. [Conditions: lean/rich switching frequency: 6/1 s; lean: 300 ppm NO, 5%  $O_2$ , balance Ar; rich: 300 ppm NO, 6.21%  $H_2$  or 0.69%  $C_3H_6$ , balance Ar, tested catalyst: PA, PCA, PBA].

occurs through its catalytic reaction with  $O_2$  due in part to unavoidable upstream mixing between the rich and lean feeds. With its limited capacity to store NOx or  $O_2$  in this temperature range, catalytic oxidation by gaseous  $O_2$  is the primary  $C_3H_6$  conversion pathway. Consumption also occurs through reaction of  $C_3H_6$  with stored NOx and/or stored oxygen. The increased conversion obtained for PBA vs. PA is probably due to its reaction with stored NOx. The much larger incremental increase for PCA is clearly a result of its reaction with oxygen stored by the ceria [29,30].

Fig. 3 shows the estimated amount of stored NOx on the three samples as a function of feed temperature. These were estimated from NOx effluent profiles by the method described in the Supplementary Material (Fig. S1). Previous NSR studies [23,44,69] show that NOx storage involves at least three sequential steps. In the first step complete NOx storage occurs, evidenced by the absence of NO or NO<sub>2</sub> in the effluent. The second step, commences at NO and/or NO<sub>2</sub> breakthrough,

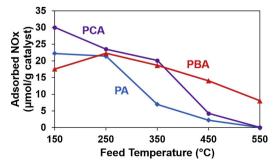


Fig. 3. Calculated amount of adsorbed NOx on PA, PCA or PBA as a function of feed temperature. [Conditions: 300 ppm NO,  $5\% \text{ O}_2$ , balance Ar].

involves an increasing effluent concentration of both species, and lasts for 10's of seconds. The third step is the much slower approach of the NOx concentration to its feed level and may last for several minutes to hours. Since the current study focuses on NOx abatement performance over a range of cycle times, the time period of interest is confined to 60 s. For temperatures below 350 °C PCA has a slightly higher NOx storage than PBA and PA. At temperatures exceeding 400 °C, PBA outperforms both PCA and PA. Even at temperatures as high as 550 °C PBA retains some NOx storage capacity while both PCA and PA have none. These data are consistent with previous works [33-35]. Bulk barium nitrate and nitrite species decompose in air at 590 °C and 267 °C, respectively [4]. In contrast, the NOx storage contributed by CeO<sub>2</sub> [33–35] and Al<sub>2</sub>O<sub>3</sub> [45.46] occurs at much lower temperatures. Ji et al. [35] reported that surface nitrates formed on Pt/CeO2 remained stable at 300 °C but disappeared upon heating to 400 °C. Al<sub>2</sub>O<sub>3</sub> may also contribute to the NOx adsorption at temperatures below 300 °C [45]. These results are under "dry" conditions - NOx storage on PA is much lower in the presence of H<sub>2</sub>O. Nitrates and nitrites formed on Al<sub>2</sub>O<sub>3</sub> with a co-feed of NO and O2 but those are weakly bound and displaced by H<sub>2</sub>O [72]. Nitrates and nitrites [4] formed on CeO<sub>2</sub> or BaO surface still survive with H<sub>2</sub>O introduction, as CeO<sub>2</sub> and BaO possess relatively stronger basicity compared to Al2O3.

A comparison between the cycle-averaged NOx conversion and NOx storage provides insight into the working mechanism. With H2 as reductant, the impact of catalyst composition on cycle-averaged NOx conversion (Fig. 2.a) follows the trends of the short-time NOx storage data (Fig. 3) with two exceptions described below. For temperatures < 300 °C, PCA has the highest NOx storage capacity (Fig. 3) and the highest NOx conversion (Fig. 2.a). In contrast, at higher temperatures ( > 400 °C), PBA exhibits the highest NOx storage (Fig. 3) and as a result outperforms PCA and PA during fast cycling (Fig. 2.a). Two exceptions to the connection between storage and conversion are encountered at 150 °C and 350 °C. At 150 °C, PBA has a higher NOx conversion but a lower NOx storage than PA. This may be explained by the difference in the stability of nitrates or nitrites on PBA and PA. That is, nitrate species formed on Al<sub>2</sub>O<sub>3</sub> can be easily purged in an inert gas at the same temperature [45] while nitrate species formed on BaO remain stable during inert purging [4]. Thus, even though PA traps more NOx during the lean feed, the loosely held NOx desorbs from the surface, preventing its reduction. At 350 °C, PCA has a lower cycle-averaged NOx conversion but has a slightly higher NOx storage compared to PBA. This contradiction may be explained by the oxygen storage of PCA. Ren et al. [33] pointed out that on a ceria-containing NSR catalyst, the reductant reacts with both stored NOx and oxygen. Thus, in cyclic experiments with PCA, additional consumption of reductants may result in a lower reductant concentration and a lower cycle-averaged NOx

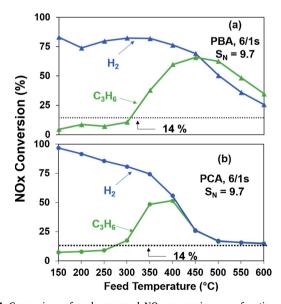
For  $C_3H_6$  as reductant, the dependence of cycle-averaged NOx conversion (Fig. 2.b) also follows the fast NOx storage trends at temperatures exceeding the propylene light-off (e.g.  $> 300\,^{\circ}$ C). With  $C_3H_6$  as reductant PCA outperforms PBA over a wider range of temperature than with  $H_2$ ; i.e., up to ~350 °C. The promotional effect of ceria is through mitigation of the self-inhibited propylene oxidation on Pt. Previous NSR studies have shown that hydrocarbons are inferior to  $H_2$  due to self-inhibition [47,48]. Lang et al. [31] showed that the addition of ceria, albeit with Pd/Al<sub>2</sub>O<sub>3</sub>, lowers the light-off temperature of  $C_3H_6$ . In summary, the consistent dependence of cycle-averaged NOx conversion on the fast NOx storage capacity validates the importance of the conventional NSR mechanism at higher cycling frequency.

Figs. 2.a and 2.b show that PCA outperforms PA over the entire feed temperature range with both  $\rm H_2$  and  $\rm C_3H_6$  as reductant. However, the difference diminishes at higher temperatures (> 400 °C). In fact, the NOx conversion approaches ~14% for both PCA and PA at feed temperatures exceeding 500 °C. This value in fact corresponds to the fraction of the cycle that is rich; i.e., 1/7, showing that the NOx reduction only occurs during the rich part of the cycle at high temperature. The

only slightly higher NOx conversion on PCA compared to PA, i.e., 4% higher for  $H_2$  and 8% for  $C_3H_6$ , suggests that the benefits of  $CeO_2$  under lean conditions ( $S_N = 9.7$ ) is limited. In our previous study [39] we reported that the rate of NO decomposition/reduction over PGM-free ceria is greatly diminished due to the higher reactivity with oxygen vacancies of  $O_2$  compared to NO. In the current study using Pt/CeO<sub>2</sub>/  $Al_2O_3$ , the NOx conversion is almost certainly confined to the rich phase when there is a large excess of  $O_2$ . These results suggest that ceria has only a limited promotional impact on the deNOx under fast cycling conditions. These conclusions are not likely to be much different if  $O_2$  is present during the reductant injection. On the other hand, there is a sustained enhancement in the NOx conversion on PBA compared to PA at high temperature. Figs. 2.a and 2.b show a NOx conversion enhancement at  $600\,^{\circ}$ C of  $^{\circ}15\%$  and  $^{\circ}28\%$  for  $H_2$  and  $C_3H_6$ , respectively.

The performance with  $Pt/CeO_2/Al_2O_3$  and  $Pt/BaO/Al_2O_3$  confirm a decidedly minor role of oxygen storage material ( $CeO_2$ ) while underscoring the critical role of NOx storage material (BaO). These findings agree with those of Reihani et al. [43] who reported that the fractional NOx conversion achieved with a Pd-Rh TWC under fast cycling was only slightly higher than 15%, the percentage of rich time among the total lean-rich cycling time for a 15% duty cycle. They attributed the inferior deNOx performance to two factors; the low NOx storage due to lack of NOx storage material, and the inferior NO oxidation due to the absence of Pt.

Figs. 4.a and. b recast the cycle-averaged NOx conversion data for PBA and PCA, respectively, to more clearly show the impact of the reductant type. At low to intermediate temperature ( $<400\,^{\circ}$ C),  $H_2$  is the superior reductant for both PBA and PCA. This finding is consistent with previously-reported results and is attributed to the much lower light-off temperature of  $H_2$  on Pt-containing catalyst [49]. At high temperature ( $>450\,^{\circ}$ C) there is no longer an advantage of  $H_2$ . In fact, to the contrary, the NOx conversion with  $C_3H_6$  slightly exceeds by ~10% that obtained with  $H_2$  on the PBA catalyst (Fig. 4a). The NOx conversion enhancement with  $C_3H_6$  under fast cycling has been previously reported [16,18–20]. Earlier studies attributed the enhancement to the formation of a thermally stable hydrocarbon surface intermediate that is selectively oxidized to  $N_2$  during the lean part of the cycle [16,24]. Most recently Ting et al. [71] show via modeling of experimental results that the advantage of  $C_3H_6$  is a result of its lower external mass



**Fig. 4.** Comparison of cycle-averaged NOx conversion as a function of feed temperature over PCA or PBA with  $\rm H_2$  or  $\rm C_3H_6$  as the sole reductant. [Conditions: lean/rich switching frequency: 6/1 s; lean: 300 ppm NO, 5%  $\rm O_2$ , balance Ar; rich: 300 ppm NO, 6.21%  $\rm H_2$  or 0.69%  $\rm C_3H_6$ , balance Ar; tested catalyst: (a) PBA; (b) PCA].

transfer rate compared to  $\rm H_2$  and the resulting lower release of NO during the rich pulse. On PCA both  $\rm H_2$  and  $\rm C_3H_6$  give essentially the same 14% NOx conversion (Fig. 4b) indicating that reduction only occurs during the rich feed. Earlier discussion pointed out that the redox pathway has only a minor contribution to the NOx reduction performance under net lean conditions ( $\rm S_N=9.7$ ). For the PCA catalyst, the nearly identical NOx conversion of ~14% above 450 °C for both  $\rm H_2$  and  $\rm C_3H_6$  suggests that any additional reduction on the ceria is negligible. We expand on the redox mechanism on ceria in the next section. Results of NH<sub>3</sub> and N<sub>2</sub>O yields over three catalysts with H<sub>2</sub> or C<sub>3</sub>H<sub>6</sub> as sole reductant are reported in the Supplementary Material (Section 2) [51–55].

#### 3.2. Role of CeO2

As introduced earlier, the enhanced NOx conversion observed during fast cycling at high temperatures, is attributed to one of three mechanisms; (1) improved NOx storage utilization, (2) adsorbed hydrocarbon intermediate pathway, or (3) ceria redox pathway. The low NOx conversion obtained during lean-rich cycling on Pt/CeO $_2$  compared to Pt/BaO suggests that the redox pathway is secondary. To investigate this further, we compare the overall and transient NOx abatement behavior over the PCA catalyst for various cycling frequencies and feed conditions. Similar to the last set of experiments, H $_2$ O or CO $_2$  were excluded from the feed to avoid complications from side reactions (e.g., water gas shift reaction) in order to isolate the impact of CeO $_2$ .

#### 3.2.1. Impact of O2

Fig. 5 describes for PCA the cycle-averaged NOx conversion (5.a, b), NH<sub>3</sub> yield (5.c, d) and N<sub>2</sub>O yield (5.e, f) as functions of feed temperature for cycle time ratios spanning 90/15 s to 6/1 s and for two feed compositions,  $S_N = 9.7$  (Figs. 5.a, c, e) and  $S_N = 1.0$  (Figs. 5.b, d, f). For comparison, results and discussion on the impact of cycling frequency over Pt/BaO/Al2O3 (PBA) are included in Supplementary Material (Section 3) [4,20,23,73]. The NOx conversion increases with decreasing cycle time over the entire feed temperature range for both lean and stoichiometric feeds. For the lean feed ( $S_N = 9.7$ ), the beneficial impact of fast cycling diminishes at temperatures > 500 °C which, is a result of less effective NO decomposition in the presence of excess O2. Hence NO is only reduced during the rich part of the cycle. In contrast, for the stoichiometric feed (S<sub>N</sub> = 1) the NOx conversion approaches a plateau that incrementally expands beyond 14% with decreasing cycle time. For example,  $X_{\text{NOx}}$   $\tilde{\ }$  34% for the 6/1 cycle. The conversion increase with decreasing cycle time at lower temperatures is a result of the improved NOx storage utilization on the ceria. It is noteworthy that the lean feed gives a higher conversion than the stoichiometric feed. This is

attributed to the higher yield of NO $_2$  from Pt-catalyzed NO oxidation, which is of positive order in O $_2$ . Crocoll et al. [64] studied the NO oxidation over Pt/Al $_2$ O $_3$  catalyst under oxygen excess conditions. It was found that the increase in the O $_2$  concentration from 1.5 vol.% to 6.0 vol.% led to a monotonic increase of NO conversion to NO $_2$  over a wide temperature range (150 °C – 500 °C). The PCA NOx storage capacity results for the S $_1$  = 1.0 and S $_2$  = 9.7 cases are shown in Fig. S3 (Supplementary Material). The results show that a higher O $_2$  feed concentration gives a higher NOx storage capacity on PCA when PCA still possesses NOx storage capacity (i.e. < 450 °C). In this case, the increased NO oxidation from the O $_2$  concentration increase results in an enhanced NOx storage, which in turn gives a higher NOx conversion under net rich conditions.

Fig. 5c and d show that the cycle-averaged NH $_3$  yield is a non-monotonic function of temperature for both feeds at all cycle times. The maximum NH $_3$  yield occurs at an intermediate temperature (250 °C–400 °C) and is slightly skewed to lower values for the lean feed. To the right of the maximum the yield approaches ~14%, indicating all the NO that reacts is converted to NH $_3$  under these conditions. The lone exception is the 6/1 cycle with the lean feed for which the yield limit is ~10%. For temperatures sufficiently left of the yield maximum, an increase in the cycle time decreases monotonically the NH $_3$  yield. This trend conveys that a more efficient oxidation accompanies the faster cycling. Previous NSR studies of Pt/BaO/Al $_2$ O $_3$  have shown that during regeneration, NH $_3$  is generated upstream from the reaction of H $_2$  with stored NOx in reaction R1:

$$Ba(NO_3)_2 + 8H_2 \rightarrow 2NH_3 + BaO + 5H_2O$$
 (R1)

The  $NH_3$  subsequently reacts with stored NOx downstream oxidized by NO and  $O_2$  to  $N_2$  by reaction R2 [50,51]:

$$3Ba(NO_3)_2 + 10NH_3 \rightarrow 8N_2 + 3BaO + 15H_2O$$
 (R2)

N<sub>2</sub>O may also form via reaction R3 [73]:

$$Ba(NO_3)_2 + 2NH_3 \rightarrow 2N_2O + BaO + 3H_2O$$
 (R3)

Lietti et al. [56] identified the different temperature thresholds for the two steps on a  $Pt/BaO/Al_2O_3$  catalyst and concluded that  $NH_3$  formation (R1) is faster than  $NH_3$  consumption (R2). This should increase the net formation of  $NH_3$  (R1). Indeed, an increasing dependence of  $NH_3$  yield with cycle frequency was observed in  $Pt/BaO/Al_2O_3$  case as shown in Fig. S2 (Supplementary Material). In contrast, with its oxygen storage capacity, PCA promotes  $NH_3$  oxidation (R2) which decreases the net generation of  $NH_3$ . The decrease in  $NH_3$  yield with shorter cycle time is attributed to enhanced oxidation reaction of  $NH_3$ .  $NH_3$  formation occurs in the front part of the monolith during the rich  $H_2$  pulse; the faster cycling increases the contact of the rich and lean feeds which results in the enhanced extent of reaction. This trend is

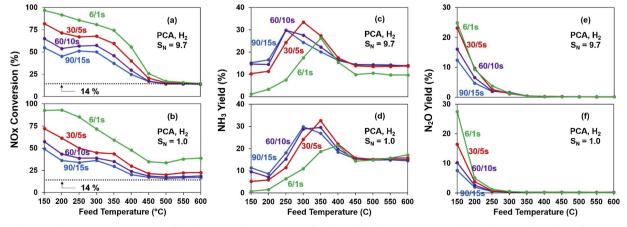


Fig. 5. Cycle-averaged NOx conversion, NH<sub>3</sub> yield and N<sub>2</sub>O yield over PCA as a function of feed temperature. [Conditions: lean/rich switching frequency:  $90/15 \, s$ ,  $60/10 \, s$ ,  $30/5 \, s$ ,  $6/1 \, s$ ; lean:  $300 \, ppm$  NO,  $5\% \, O_2 \, (S_N = 9.7)$  or  $0.5\% \, O_2 \, (S_N = 1.0)$ , balance Ar; rich:  $300 \, ppm$  NO,  $6.21\% \, H_2$ , balance Ar;].

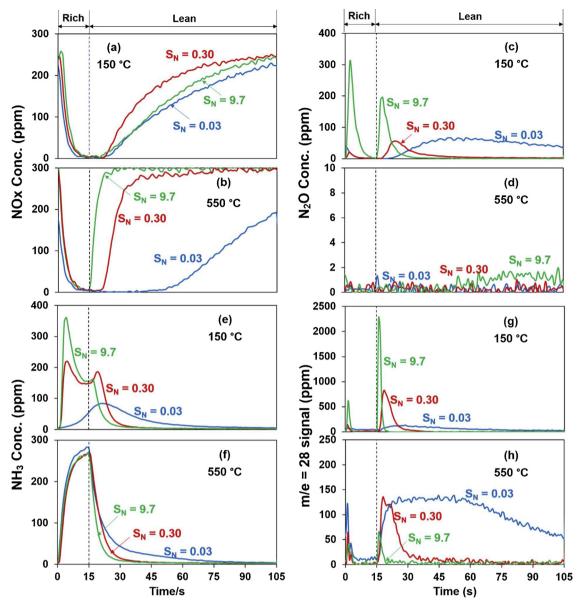


Fig. 6. Effluent Concentration profiles for a set of cycling experiments with PCA using varied  $O_2$  concentration. [Conditions: lean/rich switching frequency: 90/15 s; lean: 300 ppm NO, no  $O_2$  ( $S_N = 0.03$ ) or 1400 ppm  $O_2$  ( $S_N = 0.30$ ) or 5%  $O_2$  ( $S_N = 9.7$ ), balance Ar; rich: 300 ppm NO, 6.21%  $O_2$  H<sub>2</sub>, balance Ar; feed temperature: 150 °C or 550 °C].

consistent with Ting et al. [23] who used a NSR catalyst containing a higher loading of PGM (Pt + Rh) and mixture of BaO and  $CO_2$ . We did not include the quantitative analysis here. It is not feasible to carry out a quantitative analysis under fast cycling (e.g. 6 s lean/1 s rich) due to the FTIR gas cell mixing. The highlighted global reactions follow from the analysis of previous studies [23,51,56,66,73].

Fig. 5.e and. f show that for both feeds, the  $N_2O$  formation only occurs at low temperatures (e.g. < 300 °C) and its yield increases monotonically with decreasing cycle time for a fixed temperature.  $N_2O$  formation on PCA results from two paths; the first is a reaction between NO and reductant or surface-deposited reductive species over reduced PGM sites via the NSR pathway, while the second is reaction of NO over partially reduced ceria via the redox pathway. A shorter cycle leads to the faster oxidation/reduction transition of PGM sites and ceria, both of which promote  $N_2O$  formation and result in a higher  $N_2O$  yield.

The NOx and  $O_2$  storage functionalities and their contributions to NOx reduction may be assessed by examining the transient species concentrations at 150 °C and 550 °C. Effluent transient profiles of nitrogen-containing species provide a more detailed look at the

underlying chemistry during the lean-rich cycle (Fig. 6). The ceria NOx storage capacity decreases to a negligible value for temperatures above 400 °C [35] while the oxygen storage is sustained at high temperatures on activated ceria (i.e. > 150 °C for H<sub>2</sub>) [57]. PCA exhibits oxygen storage function at both 150 °C and 550 °C but has NOx storage function only below 450 °C. Therefore, the deNOx performance at 550 °C presents the sole contribution from redox pathway. Ting et al. [23] pointed out that mixing inside the FTIR gas cell results in a protracted tail for 20 s in this experimental setup at the given flowrate. To suppress the interference from the tail, we focus on analysis of the transient data from the longest cycling time 90/15 s. The mass spectrometer (MS) signal of m/e = 28 includes contributions from both  $N_2$  and  $N_2O$  [58] when H<sub>2</sub> is the reductant and the feed is devoid of CO<sub>2</sub>. The N<sub>2</sub> part of the m/e = 28 signal may be adjusted by subtracting the  $N_2O$  FTIR measured contribution The FTIR mixing effect prevents a complete deconvolution of the  $N_2O$  and  $N_2$  contributions from the m/e = 28 signal.

To better understand the impact of O<sub>2</sub> on NOx reduction performance over PCA, a set of experiments was conducted with varied O<sub>2</sub>

concentration in the lean feed for a 90/15 cycle at feed temperatures of 150 °C and 550 °C. Three lean feeds having different  $O_2$  concentrations (0, 1400 ppm, 5%) were applied having the following respective values of  $S_N$ : 0.03, 0.30 and  $S_N = 9.7$ . The results of these experiments are shown in Fig. 6; i.e. NOx (Fig. 6.a and b),  $N_2O$  (Fig. 6.c and d), NH<sub>3</sub> (Fig. 6.e and f), and m/e = 28 (Fig. 6.g and h). The m/e = 28 signal intensity was converted to  $N_2$  concentration using a  $N_2$  calibration factor. Following earlier comments, the  $N_2$  concentration is estimated by comparing transient plots of m/e = 28 signal and transient plots of  $N_2O$  (i.e. subtracting  $N_2O$  contribution from the overall m/e = 28 signal). Considering the limited formation of  $N_2O$  at 150 and 550 °C, Fig. 6g and h are good estimates of the  $N_2$ , notwithstanding the slower dynamics of the  $N_2O$  FTIR signal.

The rich cycling feed ( $S_N=0.03$ ) serves as a reference in that no  $O_2$  is contained in the feeds. The NOx profile exhibits a similar pattern at 150 °C (Fig. 6a) and 550 °C (Fig. 6b) despite the widely different temperatures. After a quick decrease during the rich feed, the NOx remains undetected for ~10 s and ~30 s at 150 and 550 °C, respectively. Following breakthrough, the NOx slowly increases to its feed value. At both temperatures  $N_2$  exhibits a continuous generation throughout the entire lean feed along with an additional small peak evident at the start of the rich feed. The protracted NH<sub>3</sub> tail during the lean feed exceeds the ~20 s tail produced by the FTIR cell, suggesting that NH<sub>3</sub> is generated throughout the lean period. The reduction of NO and generation of NH<sub>3</sub> in rich involves the reaction of NO with adsorbed H during the rich feed; i.e.,

$$NO + 2.5H_2 \rightarrow NH_3 + H_2O$$
 (R4)

Also, the reduction of ceria or PGM-containing ceria by  $H_2$  leads to the formation of vacancies (R5) [29,30,62] and H adatoms (R6, R7) [39], which may be utilized for NO reduction in lean period.

$$H_2 + Pt - O \rightarrow Pt + H_2O \tag{R5}$$

$$H_2 + CeO_2 \rightarrow HO - Ce - OH$$
 (R6)

$$CeO_2 + HO - Ce - OH \rightarrow Ce_2O_3 + H_2O$$
 (R7)

Reactions R8 - R10 describe NO reduction to  $N_2$  and  $NH_3$  over reduced Pt-CeO<sub>2</sub> during the lean period:

$$NO + Pt \rightarrow 0.5N_2 + O - Pt \tag{R8}$$

$$2NO + 2Ce_2O_3 \rightarrow N_2 + 4CeO_2$$
 (R9)

$$2NO + 5HO - Ce - OH \rightarrow 2NH_3 + 5CeO_2 + 2H_2O$$
 (R10)

The similar NOx,  $N_2$  and  $NH_3$  effluent profiles indicates the existence of the redox pathway.

The impact of O<sub>2</sub> on NOx reduction at high temperature (i.e. 550 °C) is discussed first due to its comparative simplicity. Previous studies [25,28] pointed out that O2 is a better oxidant than NO to re-oxidize ceria or Pt/CeO2. Correspondingly, our study shows that the addition of O2 has an inhibitory impact on NOx conversion. Fig. 6.b shows that the addition of  $O_2$  to the lean feed leads to a large increase in unreacted NOx such as an earlier and sharper slip of NOx. For the  $S_{\rm N}=9.7$  feed, NOx slips and reaches the feed level (300 ppm) almost immediately upon the switch from rich to lean feed. This suggests that NOx reduction only occurs during the rich feed and beginning of the lean feed. Accordingly, the addition of O2 strongly inhibits N2 formation during both the lean and rich feeds. This results in a smaller N<sub>2</sub> peak during the rich feed and a diminished N2 formation period during the lean feed (Fig. 6.h). For example, for the first peak  $N_2$   $\sim$  50 ppm and disappears after ~ 2 s. At 550 °C, NH<sub>3</sub> is mainly formed during the rich period with limited additional generation during the lean period. In comparison, for the  $S_N = 0.03$  case,  $NH_3$  has a large formation during the rich and as mentioned above is sustained for ~ 60 s into the lean period (Fig. 6.f). Similar to N<sub>2</sub>, the addition of O<sub>2</sub> decreases NH<sub>3</sub> generation during both the lean and rich feeds. The detrimental impact of O2 on NOx conversion at 550 °C is attributed to the competition between NO and O2 for

surface vacancies and hydroxyls.

At 150 °C the NOx reduction data reveal classical NOx storage and reduction behavior (Fig. 6.a) [33,34]. For each of the feeds, the NOx decays quickly upon the switch from lean to rich and remains undetected in the remaining part of rich feed. This trend is consistent with  $\rm H_2$  reduction of co-fed NO and NOx trapped on the ceria phase during the lean period. NOx remains undetected for ~5 s in the beginning of lean feed and then increases towards the feed concentration. The lean NOx slip is a nonmonotonic function of the  $\rm O_2$  concentration.

Previous studies [33-35] suggest that Pt/CeO2/Al2O3 exhibits considerable NOx storage capacity under net lean conditions at medium and low temperatures (< 400 °C). Haneda et al. [70] reported NO decomposition to N2 and N2O on reduced Pt/CeO2 catalysts even at room temperature. Based on previous studies [33-35,70], the transient NO trend in the lean period has the following aspects. First NOx is trapped on sites close to the Pt crystallites. This is followed by slower trapping on distal and sub-surface sites. At the same time NO decomposition occurs on reduced Pt and ceria sites. The nonmonotonic dependence of NOx reduction conversion on O<sub>2</sub> concentration at 150 °C is explained by the simultaneous involvement of two NOx reduction pathways; i.e. the conventional NSR pathway and the redox pathway. As such, for the NSR pathway, NOx is stored on the ceria during the lean feed and the NOx is reduced in the ensuing rich period. Previous studies [4,68] point out that NOx storage typically proceeds through the sequential NO oxidation to NO2 which then is reactively stored as nitrates. An alternative route involves nitrite formation. A number of studies (e.g. [63]) have shown that Pt catalyzed NO oxidation is of positive order with respect to O2. Hence, an increase in the O2 concentration should promote NOx storage on PCA. Thus, for the  $S_N = 9.7$  feed the NSR pathway has a dominant contribution to the NOx conversion as the large excess of O2 oxidizes the reduced PCA quickly and inhibits NO decomposition on PCA. At the other extreme is the  $S_{N}=0.03$  case for which the redox pathway mainly contributes to the NOx conversion as O2 is excluded in feed. As O<sub>2</sub> is added into the mixture the NO decomposition is inhibited, resulting in a lower NOx conversion. However, compared to the  $S_N=9.7$  case, the low  $O_2$  concentration in  $S_N=0.30$  case leads to a lower NO oxidation rate and NOx storage and inferior NOx reduction behavior. This explains the poor NOx abatement performance obtained for the  $S_N = 0.30$  case.

Figs. 6.c, 6.e and 6.g show the transient profiles of N-containing products at 150 °C. With the feed devoid of  $O_2$  ( $S_N = 0.03$ ), each of the N-containing products (N2, NH3 and N2O) is generated throughout the lean period, confirming the contribution from the redox pathway. The addition of O<sub>2</sub> shortens the generation period. For example, the increase of S<sub>N</sub> from 0.03-9.7 shrinks the reduction period to the first 3 s of the lean phase, with larger  $N_2$  (~ 2500 ppm) and  $N_2O$  (~200 ppm) peaks detected. Recall that O<sub>2</sub> fed during the lean period competes with NO to decompose on vacancies and/or react with residue H2 on the catalyst, an increase of O2 concentration in the lean decreases the chance for NO reduction to N2 and NH3. Therefore, the generation period of N-containing products shrinks with the addition of O2. With a large excess of  $O_2$  in feed ( $S_N = 9.7$ ), the NOx reduction pathway is dominated by the NSR pathway. The generation of N-containing species (N2, NH3, N2O) during the lean period may be attributed to the reaction between the adsorbed reducing species (e.g. NH3 and H adatoms) and NOx fed [59-61].

Transient profiles of nitrogen-containing species confirm the transition from the redox pathway to the NSR pathway with an increase in the  $\rm O_2$  concentration ( $\rm S_N$ ). The  $\rm S_N$  increase from 0.03 to 9.7 shrinks the duration during which N-containing products ( $\rm N_2$ , NH<sub>3</sub> and N<sub>2</sub>O) are generated in the rich period and early lean period. Elimination of any N<sub>2</sub>, NH<sub>3</sub> and N<sub>2</sub>O generation during the late lean phase (i.e., after 15 s in lean phase for the  $\rm S_N=9.7$  case) shows a transition from some reduction to NOx storage. Correspondingly, the expanded difference between detected NOx and the feed level (300 ppm) is an indicator of NOx storage. Also, with the increase in  $\rm S_N$  is the increased formation of

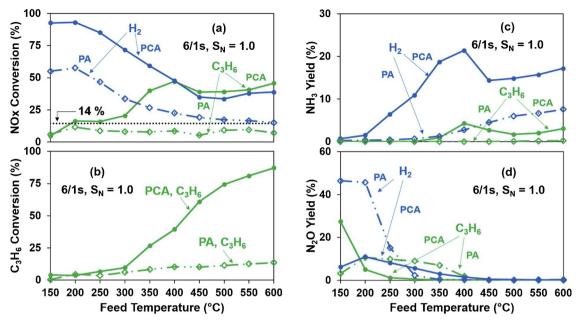


Fig. 7. Cycle-averaged reactant (NOx and C<sub>3</sub>H<sub>6</sub>) conversions and product (N<sub>2</sub>O and NH<sub>3</sub>) yields over PCA or PA as a function of feed temperature. [Conditions: lean/rich switching frequency: 6/1 s; lean: 300 ppm NO, 0.5% ppm O<sub>2</sub>, balance Ar; rich: 300 ppm NO, 6.21% H<sub>2</sub> or 0.69% C<sub>3</sub>H<sub>6</sub>, balance Ar;].

nitrogen-containing products (N2, NH3 and N2O) during the rich phase.

#### 3.2.2. Impact of reductant

The redox pathway involves NOx reduction on reduced ceria sites, including the ceria sites proximal to PGM particles and isolated ceria sites [25,28]. Previous studies show that the addition of Pt onto ceria leads to two additional active sites in addition to  $\text{CeO}_2$  sites; namely, PGM sites and Pt/ $\text{CeO}_2$  sites [29,30]. Each of these is able to reduce NO [25,28,29]. Our previous [39] study showed that ceria reduced NO once activated at high temperatures (e.g.  $> 450\,^{\circ}\text{C}$ ). To differentiate the contribution from Pt and Pt/ $\text{CeO}_2$  sites and to identify the role of Pt/ $\text{CeO}_2$  sites, PCA and PA were compared in targeted experiments described next.

We showed earlier the deNOx performance over PCA and PA for the  $S_{\rm N}=9.7\,$  case and concluded that the NOx storage capacity plays a significant role in NOx conversion. Here we focus on the  $S_{\rm N}=1.0$  case; the lower  $O_2$  concentration attenuates the NOx reduction performance contributed by the redox pathway.  $H_2$  and  $C_3H_6$  reductants are used to make our points.

Fig. 7.a shows the cycle-averaged NOx conversion for the  $S_N = 1.0$ cycle over PCA under the fastest cycling (6/1 s). Similar to the  $S_N = 9.7$ case shown earlier (Fig. 4.b), H2 outperforms C3H6 at low and intermediate temperatures (150 °C-400 °C). At high temperature the behavior is different. The NOx conversion approaches 14% for both reductants for the  $S_N = 9.7$  case. In contrast, the NOx conversion is 35–40% for both reductants for the  $S_N=1.0$  case. This indicates that some reduction occurs during the lean part of the cycle. Further, C<sub>3</sub>H<sub>6</sub> gives a slightly higher NOx conversion than H2. As discussed earlier, deNOx achieved below 400 °C is a result of NOx storage and reduction. With PCA having a negligible NOx storage capacity but nonzero OSC at higher temperatures, the NOx reduction occurring during the lean feed is a result of the redox pathway. Accordingly, for both the  $S_N=9.7$  and  $S_N = 1.0$  cases, the inferiority of  $C_3H_6$  to  $H_2$  below 400 °C is a result of differences in their reactivities on Pt. On the other hand, the superiority of  $C_3H_6$  to  $H_2$  above 450 °C for the  $S_N = 1.0$  case is its higher reactivity than H<sub>2</sub> on CeO<sub>2</sub>. Wang et al. studied the complete reduction process of CeO<sub>2</sub> [25] and Pt/CeO<sub>2</sub> [26] by H<sub>2</sub> and C<sub>3</sub>H<sub>6</sub> and concluded that C<sub>3</sub>H<sub>6</sub> leads to a higher extent of reduction on both CeO<sub>2</sub> and Pt/CeO<sub>2</sub>. They suggested that it is due to the deposition of carbonaceous species on the ceria surface. The results reported in Fig. 7 corroborate the Wang et al.

findings; i.e.  $C_3H_6$  provides more reducing sites on the  $CeO_2$  (than  $H_2$ ), either vacancies or carbonaceous material, for subsequent NO decomposition/reduction. Taking this line of reasoning one step further, we showed earlier that the deNOx attributed to the redox pathway is sensitive to the  $O_2$  concentration; i.e., competition between NO and  $O_2$  for reduced sites. The data indicate that a large excess of  $O_2$  in the  $S_N=9.7$  case results in the fast oxidation of reduced sites by  $O_2$ , largely suppressing NO conversion. In the  $S_N=1.0$  case, the  $O_2$  inhibition is somewhat less with a fraction of NO able to be reduced. As a result, an enhancement of NOx conversion by  $C_3H_6$  exists in the  $S_N=1.0$  case but not in the  $S_N=9.7$  case.

To assess the contribution from Pt to that of  $CeO_2$  or even of Pt/ $CeO_2$ , the PA sample was evaluated using the same experimental conditions. Fig. 7.a shows the NOx conversion with PA under 6/1 cycling. Similar to the  $S_N=9.7$  case (Fig. 2.a and. b), PCA is much more active than PA for NOx conversion using either  $C_3H_6$  or  $H_2$  as the reductant. Over the entire temperature range, a higher NOx conversion is achieved with  $H_2$  compared to  $C_3H_6$  for PA. In contrast, our previous study [39] showed that  $C_3H_6$  gave a better deNOx performance than  $H_2$  over  $CeO_2$  under cyclic conditions. These findings suggest that the enhancement of NOx conversion over PCA by  $C_3H_6$  above 450 °C in the  $S_N=1.0$  case is a result of contribution of the Pt/ $CeO_2$  and/or  $CeO_2$  sites. Fig. 7.b shows the  $C_3H_6$  conversion over PCA and PA samples. The higher  $C_3H_6$  conversion obtained on PCA shows the role of oxygen storage in enhancing oxidation. In the absence of ceria, a large fraction of the  $O_2$  fed during the lean leaves the reactor unreacted.

Fig. 7.c shows the NH $_3$  yield over PCA and PA with H $_2$  or C $_3$ H $_6$  as reductant. For both PCA and PA, H $_2$  is much more effective in providing H adatoms to react with NO to make NH $_3$ . For PCA the NH $_3$  yield increases with feed temperature below 400 °C and reaches a plateau above 450 °C (~ 14% with H $_2$  and ~ 3% with C $_3$ H $_6$ ). The nonmonotonic dependence of NH $_3$  yield on feed temperature for PCA may be attributed to the competing temperature dependent NOx and O $_2$  storage capacities, as we discussed earlier. The NH $_3$  yield over PA increases monotonically with feed temperature for H $_2$  feed while remaining at a negligible level for C $_3$ H $_6$ . Finally, Fig. 7.d shows the N $_2$ O yield over PCA and PA. The formation of N $_2$ O is confined to lower temperatures (< 400 °C) with PA giving a slightly higher N $_2$ O yield.

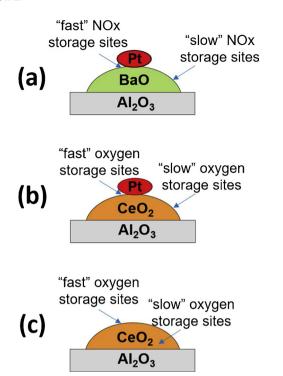


Fig. 8. Fast and slow NOx or oxygen storage sites. (a) fast and slow NOx storage sites on PBA; (b) (c) fast and slow oxygen storage sites on PCA. Here BaO and  $CeO_2$  are chosen to represent the NOx and oxygen storage material respectively.

## 3.3. Impact of fast cycling

As presented earlier, there are three principal pathways (conventional NSR pathway, HC intermediate pathway, redox pathway) that may contribute to deNOx performance during lean-rich switching. Earlier the dependence of NOx conversion and byproduct (NH3 and N2O) yield on cycling frequency for PCA catalysts was presented (Fig. 5). The results show that because PCA has NOx and O2 storage capacity, both the NSR and redox pathways contribute to extents that depend on the operating conditions. The results for PBA reveal the importance of NOx storage in enhancing NOx conversion over a wide range of temperatures (Fig. S4 in Supplementary Material). A comparison of PBA to PCA indicates that the redox pathway may only contribute at high temperature and without an excess of O2 (Fig. 2). The contribution of the hydrocarbon intermediate pathway is more difficult to assess. Recent findings indicate that it does contribute but its impact is secondary. In this section we focus on the impact of cycle frequency on the deNOx performance and explain the impact in the context of the two pathways.

It is instructive to compare and contrast the storage and utilization of NOx and O2. Fig. 8 shows schematics of the locale of an individual Pt and/or CeO<sub>2</sub> particle for the PBA and PCA catalysts. Fig. 8.a depicts the fast and slow NOx storage sites on the Pt/BaO (or Pt/CeO2 at low temperature). Previous works have described how the local structure of the NSR catalyst defines at least two populations of sites; specifically proximal and distal NOx storage sites [23,44]. Kumar et al. [65] showed experimentally that the storage and regeneration of NOx during NSR involves an active interface at the contact line of the PGM crystallite and NOx storage material. These measurements provided an experimental underpinning of a coupled process of catalytic reaction and transport. Under conditions that the NO supply is limiting, the diffusion of stored or adsorbed NOx species to and from storage sites is the rate controlling process. The process was modeled at the crystallite scale by Bhatia et al. [66] and Shakya et al. [67]. A simplified picture (and model) comprises "fast" and "slow" NOx storage sites, the former of which are proximal site and the latter distal. The primary pathway

for NOx storage involves sequential steps. First, NO and  $O_2$  adsorb and react on the Pt, forming  $NO_2$ . The  $NO_2$  then moves to a storage sites via the gas phase which involves desorption and gas phase transport, or via the surface, which involves spillover and surface or solid-state diffusion.

Fig. 8.b is a schematic picture of the Pt/CeO<sub>2</sub> catalyst which, instead of storing and consuming NOx, stores and consumes O<sub>2</sub>. The PCA catalyst comprises fast and slow oxygen storage sites based on the analogous premise of proximity of the CeO<sub>2</sub> to the PGM crystallite [29]. Fast oxygen storage sites are defined as the proximal to Pt particles whereas slow oxygen storage sites are defined as storage sites far away from Pt particles. A second pair of fast and slow oxygen storage sites can be designated according to the position on/within a CeO<sub>2</sub> particle, as per Fig. 8.c, following our previous study on NO decomposition over PGM-free ceria. The reduction of PGM-free ceria first undergoes a fast surface reduction process and then continues with slow consumption of oxygen diffusing from the bulk to the surface. As the surface oxygen is more accessible than bulk oxygen, the "fast" oxygen storage sites are the oxygen sites on the surface of ceria particles while the "slow" oxygen storage sites are the oxygen sites inside ceria particles.

For both the NSR and redox pathways, enhancement in the NOx conversion by rapid pulsing with a reductant can be ascribed as the more efficient utilization of fast sites, be they NOx storage sites or oxygen storage sites. During cyclic NSR process, the occupation and regeneration starts with the utilization of fast NOx storage sites and then proceeds with slow NOx storage sites. As pointed out in previous studies [23,43], faster switching between lean and rich feeds through a shorter cycle results in less loss of NOx. That is, a longer lean period ultimately results in the breakthrough of NO/NO<sub>2</sub> as the diffusion to sites further removed from the PGM crystallite is too slow compared to the characteristic residence time of the flowing gas. As fast NOx storage sites offers easier accessibility and faster occupation/regeneration rates, fast cycling leads to a higher NOx conversion in the NSR pathway.

A similar picture is envisioned for the redox pathway. This involves the generation of reducing components (vacancies and accumulating surface intermediates) during the rich phase feed with the subsequent adsorption and decomposition/reduction of NOx during the lean feed. Faster switching between lean and rich streams enable a more efficient utilization of the oxygen storage sites. During cyclic experiments, the occupation and regeneration first starts with the utilization of fast oxygen storage sites and then proceeds with slow oxygen storage sites. As fast oxygen storage sites offer easier accessibility and faster occupation/regeneration rates, fast cycling leads to a higher NOx conversion via the redox pathway.

When  $C_3H_6$  is the reductant, the working mechanisms are more complex due to the potential involvement of surface hydrocarbon intermediate species. Compared to  $H_2$ , the injection of  $C_3H_6$  onto both PBA and PCA results in a reduced PGM surface with accumulation of hydrocarbon species [16,26]. The adsorbed species may be further utilized for NOx reduction, either through interaction with adsorbed NOx for intermediates R-NOx and finally  $N_2$  in NSR pathway [14,16], or through interaction with NO and/or  $O_2$  for additional NO reduction in redox pathway. At high temperatures (> 550 °C), the utilization of  $C_3H_6$  results in a deeper reduction extent of ceria with accumulation of hydrocarbon intermediates [25,26]. These intermediates on serve as additional reductant and enhance NO reduction via the redox pathway.

The promotion of  $C_3H_6$  shows better resistance to  $O_2$  during NSR. Enhancement of NOx conversion by  $C_3H_6$  via a redox pathway decreases monotonically with increased  $O_2$  concentration in the feed, or increased  $S_N$  (Fig. S5 in Supplementary Material [64]), eventually becoming negligible in a large excess of  $O_2$  (e.g.  $S_N=9.7$ , Fig. 4b,  $>450\,^{\circ}\text{C}$ ). In contrast, the enhancement of NOx conversion by  $C_3H_6$  during the NSR pathway is sustained in the  $S_N=9.7$  case (Fig. 4a). The better resistance in the NSR pathway may be ascribed to the incorporation of stored NOx on storage sites (e.g. BaO). In the redox pathway, the hydrocarbon intermediates formed on PGM sites are

rapidly consumed by NO and  $O_2$  with  $O_2$  exhibiting a better oxidation capability on reduced surface. With a large excess of  $O_2$  in the feed, hydrocarbon intermediates accumulating on the PCA will be first scavenged by  $O_2$ , diminishing the likelihood of NO reduction. In contrast, over NSR catalyst (e.g. PBA), any accumulated hydrocarbon intermediates PGM are able to react directly with stored NOx, which enables enhanced NOx conversion by  $C_3H_6$  at high temperatures in a lean atmosphere.

#### 4. Conclusions

Experiments of NOx reduction under fast lean/rich cycling over Pt/  $Al_2O_3$ , Pt/CeO<sub>2</sub>/ $Al_2O_3$  and Pt/BaO/ $Al_2O_3$  provide insight into the underlying mechanism and elucidate the beneficial impacts from storage components (i.e. BaO and CeO<sub>2</sub>). The impacts of catalyst formulation, reductant type,  $O_2$  concentration, and cycling time were evaluated.

With cycle-averaged lean feeds ( $S_N = 9.7$ ) and  $H_2$  as reductant, Pt/CeO<sub>2</sub>/Al<sub>2</sub>O<sub>3</sub> (PCA) exhibits the highest NOx conversion below 300 °C while Pt/BaO/Al<sub>2</sub>O<sub>3</sub> (PBA) surpasses PCA above 300 °C. Above 450 °C, NOx reduction over PCA is largely confined to only the rich feed but PCA still surpasses Pt/Al<sub>2</sub>O<sub>3</sub> (PA) for a small enhancement in the NOx conversion. Generally, the overall NOx conversion over the three catalysts parallels the corresponding NOx storage capacity. The dependence of NOx conversion on NOx storage capacity underscores the necessity of NOx storage functionality under lean conditions. In contrast, the oxygen storage functionality only shows rather confined promotion. Pathways involved in adsorbed NOx, including the conventional NSR pathway and the hydrocarbon intermediate pathway, are responsible for the effective NOx reduction during fast cycling. The redox mechanism plays a rather minor role.

NO reduction over PCA by  $H_2$  can be achieved in both lean and rich with anaerobic feeds within the whole temperature range (150 °C ~ 600 °C). With the addition of excess  $O_2$  NOx conversion over PCA increases below 400 °C but decreases above 400 °C. The opposing trends can be attributed to the dependence of NOx storage and/or oxygen storage capacity of PCA on temperatures. PCA mainly exhibits NOx storage capacity at lower temperatures (i.e. < 400 °C) and the enhancement from excess  $O_2$  comes from the increased NOx storage rates. At higher temperatures (i.e. > 400 °C), PCA only exhibits oxygen storage capacity. The detrimental impact from excess  $O_2$  results from competition between NO and  $O_2$  for reduced sites (vacancies and adsorbed reducing intermediates).

Fast cycling enhances NOx conversion over PBA in the lean case and over PCA for the stoichiometric case within the whole temperature range using  $\rm H_2$  as the reductant. The promotional impact over PCA for the lean case only exists at low temperatures (< 450 °C), when PCA maintains NOx storage capacity. The promotional impact from fast cycling can be ascribed to the better utilization of fast NOx or oxygen storage sites from the conventional NSR pathway or redox pathway correspondingly. With a cycle-averaged rich and stoichiometric feed, lean/rich cyclic operation is inferior to steady state operation. However, lean/rich cyclic operation surpasses steady state operation with the presence of slightly excess  $\rm O_2$ .

Enhanced NOx conversion was also observed by adopting fast cycling and/or  $C_3H_6$  as reductant over PBA in the lean case or over PCA in the stoichiometric case. Compared to  $H_2$ , reduction of PBA and PCA by  $C_3H_6$  results in surface-deposited hydrocarbon fragments on Pt and/or ceria, which can be further utilized by stored NOx on PBA or gaseous fed NO in rich. The additional reduction extent of PBA and PCA provided by  $C_3H_6$  may account for the enhancement of overall NOx conversion by  $C_3H_6$ . The enhancement from  $C_3H_6$  is more resistant to  $O_2$  in NSR pathway than the redox pathway. The presence of  $H_2O$  and  $CO_2$  inhibits the enhancement by  $C_3H_6$  in both PBA and PCA cases mainly due to the competition for NOx or oxygen storage sites.

This study confirms the prevalence of the conventional NSR mechanism and the hydrocarbon intermediate pathways during fast

cycling over NSR catalysts. A redox pathway is validated but proven to extend a rather limited impact on NOx abatement under excess  $O_2$ . Fast cycling exhibits enhancement over PBA in  $S_N=9.7$  case and over PCA in  $S_N=1.0$  case, thanks to the improved utilization of fast NOx storage and oxygen storage (or vacancy) sites respectively.  $C_3H_6$  leads to an enhanced NOx conversion over PBA in the  $S_N=9.7$  case and over PCA in the  $S_N=1.0$  case. The benefits from  $C_3H_6$  may be ascribed to the deposition of hydrocarbon fragments on Pt sites, which serve as additional reduction of stored NOx or gaseous NO in NSR or redox pathway respectively.

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# Appendix A. Supplementary data

Supplementary material related to this article can be found, in the online version, at doi:https://doi.org/10.1016/j.apcatb.2019.05.044.

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